

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25

**Altered precipitation dynamics lead to a shift in herbivore dynamical regime**

Adam Pepi<sup>1,2\*</sup>, Marcel Holyoak<sup>3</sup>, Richard Karban<sup>2</sup>

<sup>1</sup>Graduate Group in Ecology, University of California Davis, Davis CA 95616

<sup>2</sup>Department of Entomology and Nematology, University of California Davis, Davis CA 95616

<sup>3</sup>Department of Environmental Science and Policy, University of California Davis, Davis CA 95616

\*Corresponding author. Email: [adampepi@gmail.com](mailto:adampepi@gmail.com)

**Keywords:** climate change, population dynamics, stochastic resonance, delayed density-dependence

**Type of article:** Letter

**Abstract word count:** 147

**Main text word count:** 3882

**Number of references:** 43

**Number of figures:** 4

**Number of text boxes:** 0

**Number of tables:** 1

**Statement of authorship:** AP, RK, and MH conceived the study, AP conducted the analyses and wrote the manuscript, RK collected the census data, and RK and MH provided editorial advice.

**Data accessibility statement:** The supporting data will be archived in HAL upon publication.

26           **Abstract**

27           The interaction between endogenous dynamics and exogenous environmental variation is  
28 central to population dynamics. Although investigations into the effects of changing mean climate is  
29 widespread, changing patterns of variation in environmental forcing also affect dynamics in complex  
30 ways. Using wavelet and time-series analyses, we identify a regime shift in the dynamics of a moth  
31 species in California from shorter to longer period oscillations over a 34-year census, and  
32 contemporaneous changes in regional precipitation dynamics. Simulations support the hypothesis that  
33 shifting precipitation dynamics drove changes in moth dynamics, possibly due to stochastic resonance  
34 with delayed density-dependence. The observed shift in climate dynamics and the interaction with  
35 endogenous dynamics mean that predicting future population dynamics will require information on  
36 both climatic shifts and their interaction with endogenous density dependence, a combination that is  
37 rarely available. Consequently, models based on historical data may be unable to predict future  
38 population dynamics.

39           **Introduction**

40           The dynamics of populations reflect the interplay between endogenous demographic and  
41 exogenous environmental drivers. Since its inception, population ecology has focused on debates about  
42 the relative contribution of these components to the generation of several salient phenomena observed  
43 in the dynamics of natural and laboratory populations, particularly cyclic fluctuations (Nicholson 1933;  
44 Andrewartha & Birch 1954; Barraquand *et al.* 2017). More recently, it has been recognized that both  
45 endogenous and exogenous drivers play important roles in generating observed population dynamics,  
46 and that endogenous deterministic dynamics and exogenous environmental noise or perturbations may  
47 combine to generate differing dynamics than would be expected for either component alone (Bjørnstad  
48 & Grenfell 2001; Turchin 2003; Barraquand *et al.* 2017). There has been additional motivation to  
49 understand the effects of climate on population dynamics as the effects of global climate change on the  
50 planet's biota have become more apparent (Walther *et al.* 2002; Parmesan 2006). Climate change is

51 expected to result in increased climate variability (Coumou & Rahmstorf 2012), as well as alterations  
52 to patterns of large-scale climate oscillations (Simon Wang *et al.* 2017), both of which are important  
53 drivers of local population dynamics. In a prominent example, changes in oceanic temperature  
54 oscillation regimes have resulted in dramatic changes to precipitation patterns in California (Simon  
55 Wang *et al.* 2017), with large impacts to society and natural ecosystems.

56 Climate variation can have direct effects on interannual fluctuations in population size or may  
57 interact in complex ways with endogenous dynamics of populations. For example, environmental  
58 perturbations can sustain population oscillations that might otherwise decay to a stable equilibrium  
59 (Tomé & De Oliveira 2009; Barraquand *et al.* 2017). Climate change has also appeared to cause the  
60 collapse of population cycles of many species across Europe (Ims *et al.* 2008; Cornulier *et al.* 2013).  
61 Environmental perturbations with different spectra of variability can also amplify, dampen, or impose  
62 their own spectra on oscillatory populations depending on the “colour” (temporal autocorrelation) of  
63 environmental spectra relative to the spectra of the endogenous dynamics of the population (Greenman  
64 & Benton 2003). Changing climate can also interact in a non-stationary way with population  
65 dynamics, such as transient effects of long-term climate oscillations on epidemic disease cycles (Rodó  
66 *et al.* 2002; Cazelles *et al.* 2005). However, testing for such effects on population dynamics of other  
67 kinds of organisms requires rarely available long-term population data.

68 In the present study, we examined how changing precipitation dynamics interact with the  
69 endogenous population dynamics of an extensively studied insect species, the Ranchman’s tiger moth  
70 (*Arctia virginalis*). We analyzed 34 years of census data from northern California, over a period during  
71 which there have been significant shifts in the dynamics of regional climate (Simon Wang *et al.* 2017).  
72 Using time series analyses and simulations, we tested for changes in population dynamics, and  
73 compared multiple possible mechanisms for observed shifts. Using simulations, we tested the  
74 hypotheses that shifting dynamics were because 1. underlying dynamics were first masked and later  
75 amplified by precipitation, 2. underlying dynamics were first amplified and later masked by

76 precipitation, or 3. dynamical shifts were driven completely exogenously by changing precipitation  
77 dynamics.

## 78 **Materials & Methods**

### 79 *Description of system*

80 Ranchman's tiger moth (*Arctia virginalis*) is a univoltine, day-active Arctiine moth, native to  
81 much of the western United States. Adult moths emerge in late spring or early summer and have a  
82 flight period of several weeks during which they do not feed. Eggs are laid on low vegetation or litter  
83 in early summer; small caterpillars hatch soon after eggs are laid. Early instar caterpillars are heavily  
84 preyed upon by ground nesting ants, and are potentially food-limited during seasonal senescence of  
85 vegetation in Mediterranean summers (Karban *et al.* 2013, 2017). Time series analyses have shown  
86 that greater precipitation during the previous year results in greater population growth, possibly due to  
87 increased food availability during the summer drought (Karban & de Valpine 2010; Karban *et al.*  
88 2017). Caterpillars feed continuously over the winter period and do not diapause in California. After  
89 this, they move up to feed on higher vegetation and become more visible in late winter in California.  
90 Caterpillars are generalists, with a preference for alkaloid-containing hosts (English-Loeb *et al.* 1993;  
91 Karban *et al.* 2010). Caterpillars are also frequently attacked by tachinid parasitoids, *Thelaira*  
92 *americana* (Karban & de Valpine 2010), which are specialists on Arctiine moths (Arnaud 1978), and  
93 may functionally be specialists on Ranchman's tiger moth at our study site, the Bodega Marine  
94 Reserve. However, analyses have suggested that parasitism has little effect on caterpillar population  
95 dynamics (Karban & de Valpine 2010). Caterpillar populations at the Bodega Marine Reserve and  
96 other sites often exhibit high mortality rates after high population density years due to a granulovirus.  
97 Monitoring at Bodega and other sites has shown delayed density-dependent infection and mortality  
98 rates due to granulovirus (Pepi, Pan & Karban, unpublished data).

### 99 *Censuses*

100 Caterpillar censuses were conducted on perennial evergreen yellow bush lupine (*Lupinus*  
101 *arboreus*) bushes at Bodega Marine Reserve in Sonoma County, California (38°19'05"N,  
102 123°04'12"W). The number of caterpillars on 10 lupine bushes in the same patch were counted yearly  
103 from 1986 to 2019 (>10 in 1986). Bushes were censused in the last week of March each year. Previous  
104 repeated censuses within a year between late February and the end of March suggest that population  
105 estimates were unlikely to vary significantly due to changes in seasonal phenology from year to year  
106 (Karban and Grof-Tisza, unpublished data), because of the long development period of caterpillars and  
107 their limited mobility during this stage. The same lupine bushes were censused each year; however, the  
108 identity of bushes changed because these lupines were short lived (< 7yr). To account for variation in  
109 sampling effort, area of each lupine bush censused was measured to calculate caterpillar density per m<sup>2</sup>  
110 (plotted as caterpillars per 100 m<sup>2</sup> in Figure 3 for legibility). Precipitation was recorded at the site as  
111 part of ongoing climate monitoring by the University of California and using a rain gauge at the study  
112 site (US Weather Bureau type manual rain gauge prior to 1992 and an optical rain gauge ORG-815,  
113 Optical Scientific, Gaithersburg, MD since 1992 with a Hydrological Services TB4 tipping bucket,  
114 Campbell Scientific, Ogden, UT since 2003). For analyses, total annual precipitation within the  
115 hydrologic year was calculated (from October 1<sup>st</sup> of the previous year to September 30<sup>th</sup> of the current  
116 year).

### 117 *Statistical analyses*

118 To test for non-stationary relationships between precipitation and caterpillar dynamics over  
119 time, wavelet analysis was conducted separately on logged and scaled caterpillar and scaled  
120 precipitation time series, and as wavelet coherence analyses on both series (Figure 3). Scaling was  
121 accomplished by subtracting the mean and dividing by the standard deviation (scale() in R). Analyses  
122 were conducted using the package BIWAVELET (Gouhier *et al.* 2019), using Morlet wavelet transforms.  
123 In addition, change-point analyses were conducted using the SEGLM and TSDYN packages (Antonio &  
124 Stigler 2009; Stigler 2019), with a model containing direct and delayed density-dependence, and a

125 separate model also including precipitation as a covariate. We included precipitation as a covariate  
 126 based on previous knowledge that precipitation was important to dynamics. We included direct density-  
 127 dependence based on previous detection in time series analyses (Karban & de Valpine 2010), and  
 128 delayed density-dependence based on wavelet periodogram results and the observation of delayed-  
 129 density dependent mortality from granulovirus in field studies (Pepi, Pan & Karban, unpublished data).  
 130 Models with and without a threshold ( $C$ ) were compared using AIC. Break point models were of the  
 131 form:

$$132 \quad X_t \sim \text{Normal}(a_{0,1} + a_{1,1}X_{t-1} + a_{2,1}X_{t-2} + \beta_{1,1}\text{Precip}_{t-1}, \sigma_1^2) \mid t \leq C$$

$$133 \quad X_t \sim \text{Normal}(a_{0,2} + a_{1,2}X_{t-1} + a_{2,2}X_{t-2} + \beta_{1,2}\text{Precip}_{t-1}, \sigma_2^2) \mid t \geq C,$$

134 in which  $X_t$  is log population density [ $\ln(\text{count}/\text{area})$ ],  $a_0$  is the intercept,  $a_1$  is direct density  
 135 dependence,  $a_2$  is delayed density dependence, and  $\beta_1$  is the effect of precipitation,  $\sigma^2$  is the variance,  
 136 with a separate parameter estimate for each before and after the threshold.

137 Bayesian state space population models using a Poisson observation process were constructed  
 138 to test for direct and delayed density-dependence and effects of precipitation. We conducted this as a  
 139 separate step from testing for thresholds to avoid identifiability issues due to limited data availability.  
 140 We primarily examined a model with the same process structure as breakpoint models, including direct  
 141 and delayed density dependence and an effect of precipitation, based on *a priori* knowledge about the  
 142 system. For comparison, models with all possible combinations of variables were generated and  
 143 compared using WAIC (Vehtari et al. 2017). State space models were fit to the time series from 1986-  
 144 2004 and 2004-2019 separately and results compared, based on findings of change-point analyses. The  
 145 full state-space model was of the form:

$$146 \quad Y_t \sim \text{Poisson}(\exp(X_t) * \text{area}_t)$$

$$147 \quad X_t \sim \text{Normal}(a_0 + a_1X_{t-1} + a_2X_{t-2} + \beta_1\text{Precip}_{t-1}, \sigma^2),$$

148 in which  $Y_t$  is caterpillar count,  $X_t$  is the estimated population density state on a log scale,  $area_t$  is the  
149 area of lupine sampled,  $a_0$  is the intercept,  $a_1$  is direct density dependence,  $a_2$  is delayed density  
150 dependence,  $\beta_1$  is the effect of precipitation, and  $\sigma^2$  is the process variance. Models were fitted in JAGS  
151 with interface in R using RJAGS (Plummer 2019) and R2JAGS (Yu-Sung Su & Yajima 2015). We used a  
152 vague regularizing Gaussian prior for all parameters [ $Normal(0, 10)$ ], except for the process variance  
153 ( $\sigma^2$ ), for which we used a uniform prior [ $Uniform(0, 10)$ ]. Model convergence was assessed using  
154 the CODA (Martyn *et al.* 2019) package, by visualizing chains (Figures S1-3) and the  $\hat{R}$  convergence  
155 criterion (Gelman & Rubin 1992). We also conducted one-step ahead simulations as posterior  
156 predictive checks (Figures S4-6), for which P-values were  $0.5 \pm 0.03$ , indicating acceptable model fit.  
157 Models were fitted using 3 MCMC chains of 20,000 iterations, with 1,000 iterations of burn-in. All  $\hat{R}$   
158 values were  $< 1.001$ .

159 To ascertain the mechanisms driving shifts in dynamics, we conducted deterministic simulations  
160 by projecting populations into the future using parameter values sampled from the posteriors of fitted  
161 state space process models. Simulations were also conducted with fitted density-dependence  
162 parameters, but with all precipitation effects drawn from the posterior of the model with the highest  
163 estimated effect of precipitation (from the second half of the series;  $\beta_1 = 0.922$ ). Another simulation was  
164 conducted using the fitted process model from the second part of the series, but with density-dependent  
165 parameters ( $a_1, a_2$ ) set to zero. For simulations, observed starting population sizes were used, and  
166 observed precipitation values were used for the entire period. For each mechanistic scenario, 10,000  
167 simulations were conducted, each based on a separate draw from posteriors. Simulated population  
168 trajectories were wavelet transformed, and a dissimilarity relative to the true population series was  
169 calculated based on the method of Rouyer *et al.* (2008b) (Table S3, Figures S7-8), all using BIWAVELET  
170 (Gouhier *et al.* 2019).

## 171 Results

172 Caterpillar population dynamics exhibited a clear regime shift during our study. Wavelet  
173 spectrograms show that dominant oscillatory periods of caterpillar and precipitation dynamics shifted  
174 from short-period (2-3 yr) to long-period oscillations (4-6 yr; Figure 1c,e), though periodicity was only  
175 significant at the 95% level for precipitation in the first part of the series (2-3 yr periodicity from  
176 ~1992-1999) and caterpillars in the second part of the series (4-6 yr periodicity from ~2003-2013).  
177 Precipitation dynamics changed after ca. 1999, and caterpillar dynamics changed shortly thereafter (ca.  
178 2002). Wavelet coherence between precipitation and caterpillar numbers shifted from a 3-year period in  
179 the early part of the series to a 3-6 year period after ca. 2005 (Figure 1g), suggesting a role of  
180 precipitation in shifts in caterpillar population dynamics. The observed shift in precipitation dynamics  
181 in turn was likely caused by shifting oceanic climate oscillations; the Pacific Decadal Oscillation and  
182 offshore sea surface temperature switched from a warm to a cold phase after 1999, which resulted in  
183 shifts in dynamics of several marine species at that time (Cloern *et al.* 2010; Thomson *et al.* 2010). A  
184 similar climate regime of high-amplitude, long-period oscillations between multi-year drought and high  
185 precipitation is expected to be the norm for California in the future (Swain *et al.* 2018).

186 Change-point analyses found a change in dynamics with a threshold in 2002 ( $\Delta AIC=4.8$  relative  
187 to model without a threshold) in a model without precipitation, or in 2004 in a model including  
188 precipitation ( $\Delta AIC=6.5$ ). Before the threshold, direct density-dependence was estimated to be negative  
189 and this became positive after the threshold, though there was limited evidence that these estimates  
190 were different from zero ( $0.10 > P > 0.09$ ; see Table 1). In models that included precipitation, its effects  
191 were always near zero before the threshold, and strongly positive after the threshold (Table 1). Delayed  
192 density-dependence had negative parameter estimates in all models, with the most evidence for delayed  
193 density dependence after the threshold in the model without precipitation ( $P=0.019$ ), and weaker  
194 evidence otherwise ( $P > 0.2$ ).

195 Results from Bayesian Poisson state-space models corresponded broadly with those from the  
196 change point analyses, showing a shift in dynamics from the first to the second period. The full model



197 was an acceptable fit relative to other model structures tested, though large differences were not  
198 detected due to the state-space structure and limited data availability (Table S1;  $\Delta\text{WAIC}<2$  for all  
199 models). 90% high density posterior intervals (HDPI) that are superior for characterizing MCMC  
200 posteriors (because of more samples in the tails; Kruschke 2014) show the following: Considering the  
201 full model fitted to the entire series (1986-2019), direct density-dependence was weak ( $a_1 = -0.11$ ; 90%  
202 HDPI:  $-0.23 - 0.47$ ), delayed density-dependence was negative ( $a_2 = -0.3$ ; 90% HDPI:  $-0.63 - 0.03$ ),  
203 and there was a weak positive effect of rainfall ( $\beta_1 = 0.30$ ; 90% HDPI:  $-0.08 - 0.70$ ; Figure 2). For the  
204 first part of the series (1986-2004), direct density-dependence was negative ( $a_1 = -0.63$ ; 90% HDPI:  $-$   
205  $1.19 - -0.10$ ), delayed density-dependence was negative but weak ( $a_2 = -0.27$ ; 90% HDPI:  $-0.78 -$   
206  $0.26$ ), and the effect of precipitation was weak ( $\beta_1 = 0.26$ ; 90% HDPI:  $-0.17 - 0.68$ ; Figure 2). For the  
207 second part of the series (2004-2019), direct density-dependence was positive ( $a_1 = 0.46$ ; 90% HDPI:  
208  $0.03 - 0.82$ ), delayed density-dependence was negative but weak ( $a_2 = -0.19$ ; 90% HDPI:  $-0.63 -$   
209  $0.36$ ), and the effect of precipitation was stronger and positive ( $\beta_1 = 0.92$ ; 90% HDPI:  $0.2 - 1.75$ ;  
210 Figure 2). Overall, the results of these models provide evidence for a shift from negative to positive  
211 direct density-dependence from the first to the second part of the series, with non-overlapping 90%  
212 intervals between the two parts ( $a_1 = -1.19 - -0.13$  vs.  $0.03 - 0.82$ ), although 95% intervals did have a  
213 marginal overlap ( $a_1 = -1.29 - 0.03$  vs.  $-0.05 - 0.96$ ; Figure 2). This corresponds with a shift from type  
214 III to type IV dynamics (Figure 3) and a shift from shorter to longer period dynamics.

215 The results from simulation analyses suggested that the long-period dynamics observed in the  
216 second part of the series (2004-2019) could best recreate the observed dynamics. Specifically, there  
217 was some evidence that endogenous dynamics as parameterized from the second part of the series best  
218 recovered the observed shift in dynamical regime when used to simulate dynamics for the entire series,  
219 based on the maximum *a posteriori* dissimilarity calculated from wavelet transforms (Figure 4,  
220 S1, Table S3; Rouyer *et al.* 2008; Gouhier *et al.* 2019). This was the case both in simulations which

221 used the original fitted precipitation parameters, and in simulations in which the effect of precipitation  
222 was set to the same value to examine solely the effect of different endogenous dynamics (lowest  
223 dissimilarity:  $d=17.6$ ; whole series:  $d=22.1, 18.5$ ; first part:  $d=22.8, 21.4$ ; Figure 4, S1, Table S3).  
224 Simulations including density-dependence reproduced dynamics somewhat better ( $d=17.6$ ) than  
225 simulations without density-dependence ( $d=19.7$ ). The broad posterior intervals of simulation results  
226 indicated substantial uncertainty as to the mechanisms driving observed shifts (Figure 4). However, to  
227 the extent that we are interested in the question of whether specific endogenous dynamics (i.e., specific  
228 parameter values) underlie the observed shift in dynamics, the point estimates from the second part of  
229 the series best reproduced dynamics. Overall, results were consistent with the interpretation that the  
230 shift in dynamics was driven either by an interaction between endogenous dynamics and precipitation,  
231 or possibly solely by precipitation (Figure 4).

## 232 **Discussion**

233 Our analyses together suggest that over the census period, the changing structure of variation in  
234 precipitation dynamics interacted with the structure of endogenous dynamics of caterpillar populations  
235 to generate novel dynamics. This resulted in higher amplitude, long-period oscillations in the second  
236 part of the series (2004-2019), in which both the lowest (2005) and the highest (2019) caterpillar  
237 population densities were observed. This is in contrast with the first part of the census (1986-2004)  
238 which was characterized by weak lower amplitude and short-period oscillations. These shifts in  
239 oscillatory period corresponded with a shift from negative direct and delayed density-dependence (type  
240 III dynamics) to positive direct and negative delayed density dependence (type IV dynamics). Although  
241 many parameter posterior intervals overlapped zero, there was strong statistical evidence for our main  
242 hypothesis, showing a shift from direct density-dependence to negative density dependence during the  
243 study (Figure 2,3). This shift in dynamics appears to have been due to changing patterns of variation in  
244 environmental forcing and illustrates the complexity of forecasting impacts of changes in both mean  
245 and pattern of variation in future climates on population dynamics.

246 Simulation studies have shown that environmental noise can resonate with the dominant period  
247 of deterministic endogenous dynamics of a system if the noise spectra includes the period of the  
248 deterministic system (Royama 1992; Alonso *et al.* 2007). Environmental variation may also impose its  
249 own spectral signature on population dynamics with different dominant periods (Greenman & Benton  
250 2003). Given these observations, there are multiple possible mechanistic explanations for the shift in  
251 dynamics observed in this study. One possible interpretation is that the interaction between the  
252 endogenous dynamical structure of this population with changing exogenous perturbation (*i.e.*,  
253 precipitation) obscured the endogenous dynamics in one part of the series, but not the other (Ranta *et*  
254 *al.* 2000). Another interpretation is that dynamical shifts may have been entirely externally forced by  
255 changing precipitation dynamics. We distinguished between these possibilities by simulating  
256 deterministic population trajectories from fitted state-space models using observed precipitation, which  
257 indicated that the dynamics from the second part of the series were somewhat more likely to have  
258 represented the underlying endogenous dynamics of the system. This indicates that short-period  
259 oscillations in precipitation may have interfered with delayed density-dependence in the endogenous  
260 dynamics to generate the observed population dynamics in the first half of the caterpillar time series.  
261 By imposing short-period oscillations onto population dynamics, external forcing by precipitation may  
262 have created only the appearance of negative direct density-dependence. During the second half of the  
263 series, longer period oscillations of precipitation may have resonated with delayed density-dependence  
264 and generated high-amplitude long-period oscillations. This interpretation is supported by the fact that  
265 simulations including negative direct density-dependence (*i.e.*, the model from the first part of the  
266 series) prevented the resonance of precipitation with delayed density-dependence and did not recreate  
267 the original dynamics quite as effectively as models with positive direct density-dependence (Figure 4,  
268 S7-8, Table S3). Furthermore, the possibility that observed shifts in dynamics were driven completely  
269 externally by precipitation seems less likely because simulations lacking density-dependence did not as  
270 effectively recover the original shift in dynamics as simulations including density-dependence.

271 However, if we consider the uncertainty of parameter estimates in the simulation results, we cannot be  
272 confident in distinguishing between alternative scenarios, as all simulation posteriors overlap  
273 substantially (Figure 4). Nonetheless, despite the limited information contained within our 34-year  
274 population time series, our simulations allowed us to compare the relative evidence for alternate  
275 mechanisms that might have caused the observed shift.

276         The dynamics that we describe during the first part of the series are consistent with previous  
277 work, which indicated that precipitation positively affects caterpillar population growth rates and  
278 interacts with endogenous dynamics of overcompensating negative direct density-dependence (Karban  
279 & de Valpine 2010; Karban *et al.* 2017). Mechanisms proposed to explain the effect of precipitation  
280 include limited resources during summer drought (Karban & de Valpine 2010), or negative effects of  
281 precipitation on predatory ants (Karban *et al.* 2017). Parasitism and viral infection are potential  
282 mechanisms for both direct and delayed density-dependence in this population, as they are in many  
283 insects and particularly Lepidoptera (Myers & Cory 2013, 2016). Parasitism and viral infection can  
284 induce density-dependence in insect population dynamics when parasitoids are host-specific and  
285 display numerical responses to host density, and when virus transmission depends on host density  
286 (Myers & Cory 2013). Delays in the action of density-dependence may be caused by delayed numerical  
287 responses of parasitoids (Myers & Cory 2013), or a greater prevalence of covert viral infections or viral  
288 occlusion bodies in the environment after high-density years (Myers & Cory 2016). Long-term  
289 monitoring data revealed no delayed density-dependent parasitism in this population (Karban & de  
290 Valpine 2010). However, laboratory rearing of *A. virginalis* from multiple monitored populations  
291 suggested a delayed-density dependent rate of viral infection (Pepi, Pan & Karban, unpublished data)  
292 as is the case in many Lepidoptera (Anderson & May 1980; Myers & Cory 2013, 2016).

293         In contrast, the dynamics of the second part of the series (2004-2019) were not predictable from  
294 our previous understanding of caterpillar population dynamics derived from analyses of time series that  
295 were long by ecological standards (20 years, 1986-2006; Karban & de Valpine 2010). This type of

296 non-stationarity due to shifts in climatic regimes has large implications for forecasting and managing  
297 populations of threatened or pest species, because such shifts have the potential to obfuscate  
298 predictions about management actions. Ecological forecasting has become an urgent goal in light of  
299 global change and unprecedented human pressures on the biosphere (Clark *et al.* 2001). Whereas most  
300 literature has focused on predicting ecological state variables (e.g., population size; Dietze 2017), we  
301 examined how changing patterns of climate interacted with endogenous population drivers to  
302 qualitatively change dynamics. This illustrates the importance of considering the interaction between  
303 endogenous population drivers and exogenous climate variation in projecting population dynamics into  
304 the future, and argues that incorporating changes in patterns of climate into predictions is essential.

305         The fact that the qualitative range of dynamics in populations is to some extent limited (i.e.,  
306 there are not ten million types of population dynamics; Lawton 1992) makes predicting shifts in  
307 population dynamics due to climate change a more attainable prospect. Consistent with this, most  
308 populations have either first or second order, and either chaotic or non-chaotic dynamics (Types I-IV  
309 and I'-IV' in Fig. 2; also see Royama 1992), in addition to some other important axes of variation  
310 (Turchin 2003; Barraquand *et al.* 2017). Ecologically, the presence of overcompensating (type II-III)  
311 density dependence, a stable equilibrium (type I), or longer period cycles (type IV) have important  
312 effects on species interactions, ecosystem dynamics, and how climate is likely to affect dynamics  
313 (Ranta *et al.* 2000; Ims *et al.* 2008). Mechanistic studies separating endogenous from exogenous  
314 components of dynamics can distinguish whether observed dynamics, such as cycles, arise from  
315 different mechanisms. Some of these mechanisms include self-sustaining or noise-sustained second  
316 order dynamics, externally-forced first order dynamics (Barraquand *et al.* 2017), and non-cyclic  
317 dynamics that mask endogenous second-order dynamics, as we found support for in the present study  
318 (Greenman & Benton 2003). The application of methods such as those implemented in the present  
319 study can help distinguish between different possible combinations of endogenous and exogenous  
320 components of a system that might have generated the observed dynamics. Doing so will improve our

321 ability to understand how changes in exogenous forcing due to climate change are likely to affect future  
322 population dynamics.

### 323 **Acknowledgments**

324 Many former students of RK contributed to census data collection, and this work would not  
325 have been possible without them. These censuses were conducted at the UC Bodega Marine Reserve  
326 and we thank Peter Connors and Jackie Sones for facilitating our work there. We would also like to  
327 thank Eric Post, Jay Rosenheim, Rolf Ims, and Louis Botsford for help improving the manuscript. This  
328 work has been supported by NSF, most recently NSF-LTREB-1456225

### 329 **Data Accessibility**

330 Data will be archived on HAL upon acceptance.

### 331 **References**

332 Alonso, D., McKane, A.J. & Pascual, M. (2007). Stochastic amplification in epidemics. *J. R. Soc.*  
333 *Interface*, 4, 575–582.

334 Anderson, R.M. & May, R.M. (1980). Infectious Diseases and Population Cycles of Forest Insects.  
335 *Science* (80-. ), 210, 658–661.

336 Andrewartha, H.G. & Birch, L.C. (1954). *The distribution and abundance of animals*. University of  
337 Chicago Press.

338 Antonio, F.D.N. & Stigler, J.L.A.M. (2009). tsDyn: Time series analysis based on dynamical systems  
339 theory.

340 Arnaud, P.H. (1978). *Host Parasite Catalog of North American Tachinidae (Diptera)*. Department of  
341 Agriculture, Science and Education Administration.

342 Barraquand, F., Louca, S., Abbott, K.C., Cobbold, C.A., Cordoleani, F., DeAngelis, D.L., *et al.* (2017).  
343 Moving forward in circles: challenges and opportunities in modelling population cycles. *Ecol.*  
344 *Lett.*, 20, 1074–1092.

345 Bjørnstad, O.N. & Grenfell, B.T. (2001). Noisy clockwork: time series analysis of population

- 346       fluctuations in animals. *Science*, 293, 638–43.
- 347   Cazelles, B., Chavez, M., McMichael, A.J. & Hales, S. (2005). Nonstationary influence of El Niño on  
348       the synchronous dengue epidemics in Thailand. *PLoS Med.*, 2, 0313–0318.
- 349   Clark, J.S., Carpenter, S.R., Barber, M., Collins, S., Dobson, A., Foley, J.A., *et al.* (2001). Ecological  
350       forecasts: an emerging imperative. *Science* (80-. ), 293, 657–660.
- 351   Cloern, J.E., Hieb, K.A., Jacobson, T., Sans, B., Di Lorenzo, E., Stacey, M.T., *et al.* (2010). Biological  
352       communities in San Francisco Bay track large-scale climate forcing over the North Pacific.  
353       *Geophys. Res. Lett.*, 37, 1–6.
- 354   Cornulier, T., Yoccoz, N.G., Bretagnolle, V., Brommer, J.E., Butet, A., Ecke, F., *et al.* (2013). Europe-  
355       wide dampening of population cycles in keystone herbivores. *Science* (80-. ), 340, 63–66.
- 356   Coumou, D. & Rahmstorf, S. (2012). A decade of weather extremes. *Nat. Clim. Chang.*, 2, 491–496.
- 357   Dietze, M.C. (2017). Prediction in ecology: a first-principles framework. *Ecol. Appl.*, 27, 2048–2060.
- 358   English-Loeb, G.M., Brody, A.K. & Karban, R. (1993). Host-plant-mediated interactions between a  
359       generalist folivore and its tachinid parasitoid. *J. Anim. Ecol.*, 62, 465.
- 360   Gelman, A. & Rubin, D.B. (1992). Inference from iterative simulation using multiple sequences. *Stat.*  
361       *Sci.*, 7, 457–472.
- 362   Gouhier, T.C., Grinsted, A. & Simko, V. (2019). R package biwavelet: Conduct Univariate and  
363       Bivariate Wavelet Analyses.
- 364   Greenman, J. V. & Benton, T.G. (2003). The amplification of environmental noise in population  
365       models: Causes and consequences. *Am. Nat.*, 161, 225–239.
- 366   Ims, R.A., Henden, J.A. & Killengreen, S.T. (2008). Collapsing population cycles. *Trends Ecol. Evol.*,  
367       23, 79–86.
- 368   Karban, R., Grof-Tisza, P. & Holyoak, M. (2017). Wet years have more caterpillars: Interacting roles of  
369       plant litter and predation by ants. *Ecology*.
- 370   Karban, R., Karban, C., Huntzinger, M., Pearse, I. & Crutsinger, G. (2010). Diet mixing enhances the

- 371 performance of a generalist caterpillar, *Platyprepia virginalis*. *Ecol. Entomol.*, 35, 92–99.
- 372 Karban, R., Mata, T.M., Grof-Tisza, P., Crutsinger, G. & Holyoak, M.A. (2013). Non-trophic effects of  
373 litter reduce ant predation and determine caterpillar survival and distribution. *Oikos*, 122, 1362–  
374 1370.
- 375 Karban, R. & de Valpine, P. (2010). Population dynamics of an Arctiid caterpillar-tachinid parasitoid  
376 system using state-space models. *J. Anim. Ecol.*, 79, 650–661.
- 377 Kruschke, J. (2014). Doing Bayesian data analysis: A tutorial with R, JAGS, and Stan.
- 378 Lawton, J.H. (1992). There are not 10 million kinds of population dynamics. *Oikos*, 63, 337–338.
- 379 Martyn, A., Best, N., Cowles, K., Vines, K., Bates, D., Almond, R., *et al.* (2019). Package ‘coda’:  
380 Output Analysis and Diagnostics for MCMC.
- 381 Myers, J.H. & Cory, J.S. (2013). Population Cycles in Forest Lepidoptera Revisited. *Annu. Rev. Ecol.*  
382 *Evol. Syst.*, 44, 565–592.
- 383 Myers, J.H. & Cory, J.S. (2016). Ecology and evolution of pathogens in natural populations of  
384 Lepidoptera. *Evol. Appl.*, 9, 231–247.
- 385 Nicholson, A.J. (1933). The balance of animal populations. *J. Anim. Ecol.*, 2, 132 – 178.
- 386 Parmesan, C. (2006). Ecological and evolutionary responses to recent climate change. *Annu. Rev. Ecol.*  
387 *Evol. Syst.*, 37, 637–669.
- 388 Plummer, M. (2019). rjags: Bayesian Graphical Models using MCMC.
- 389 Ranta, E., Lundberg, P., Kaitala, V. & Laakso, J. (2000). Visibility of the environmental noise  
390 modulating population dynamics. *Proc. R. Soc. B Biol. Sci.*, 267, 1851–1856.
- 391 Rodó, X., Pascual, M., Fuchs, G. & Faruque, A.S.G. (2002). ENSO and cholera: A nonstationary link  
392 related to climate change? *Proc. Natl. Acad. Sci. U. S. A.*, 99, 12901–12906.
- 393 Rouyer, T., Fromentin, J.-M., Stenseth, N.C. & Cazelles, B. (2008). Analysing multiple time series and  
394 extending significance testing in wavelet analysis. *Mar. Ecol. Prog. Ser.*, 359, 11–23.
- 395 Royama, T. (1992). *Analytical Population Dynamics*. Chapman & Hall, London.



- 396 Simon Wang, S.Y., Yoon, J.H., Becker, E. & Gillies, R. (2017). California from drought to deluge. *Nat.*  
397 *Clim. Chang.*, 7, 465–468.
- 398 Stigler, M. (2019). seglm: Segmented/threshold regression methods.
- 399 Swain, D.L., Langenbrunner, B., Neelin, J.D. & Hall, A. (2018). Increasing precipitation volatility in  
400 twenty-first-century California. *Nat. Clim. Chang.*, 8, 427–433.
- 401 Thomson, J.R., Kimmerer, W.J., Brown, L.R., Newman, K.B., Mac Nally, R., Bennett, W.A., *et al.*  
402 (2010). Bayesian change point analysis of abundance trends for pelagic fishes in the upper San  
403 Francisco Estuary. *Ecol. Appl.*, 20, 1431–1448.
- 404 Tomé, T. & De Oliveira, M.J. (2009). Role of noise in population dynamics cycles. *Phys. Rev. E - Stat.*  
405 *Nonlinear, Soft Matter Phys.*, 79, 1–8.
- 406 Turchin, P. (2003). *Complex Population Dynamics: A Theoretical/Empirical Synthesis*. Princeton  
407 University Press, Princeton.
- 408 Vehtari, A., Gelman, A. & Gabry, J. (2017). Practical Bayesian model evaluation using leave-one-out  
409 cross-validation and WAIC. *Stat. Comput.*, 27, 1413–1432.
- 410 Walther, G., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T.J.C., *et al.* (2002). Ecological  
411 responses to recent climate change. *Nature*, 416, 389–395.
- 412 Yu-Sung Su & Yajima, M. (2015). R2jags: Using R to Run “JAGS.”

413

414

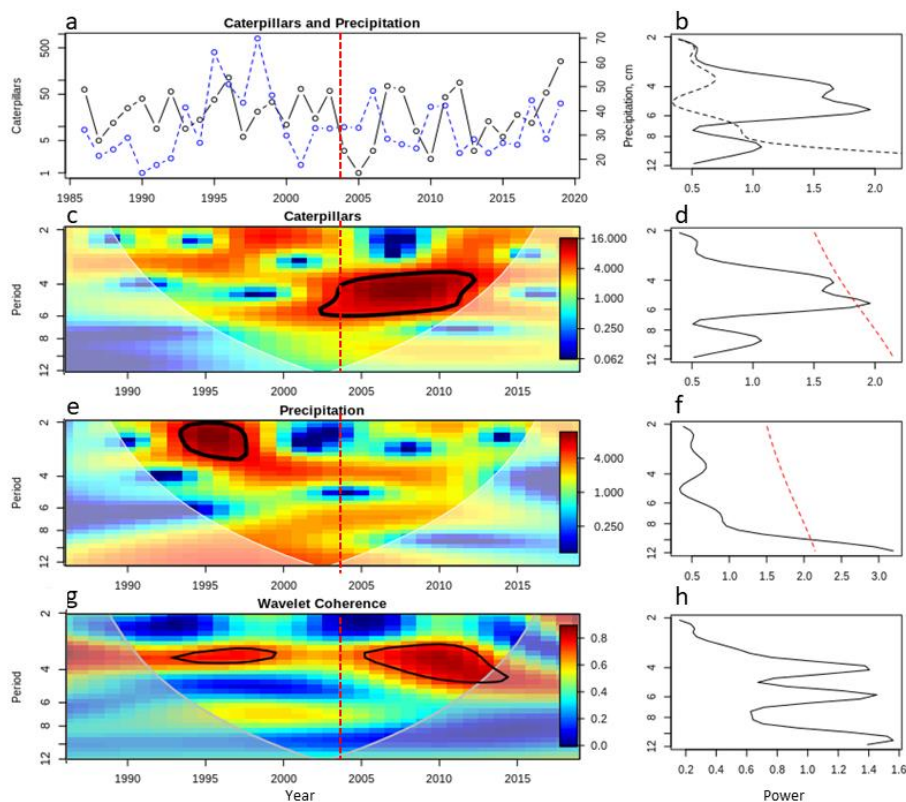
415

416

417

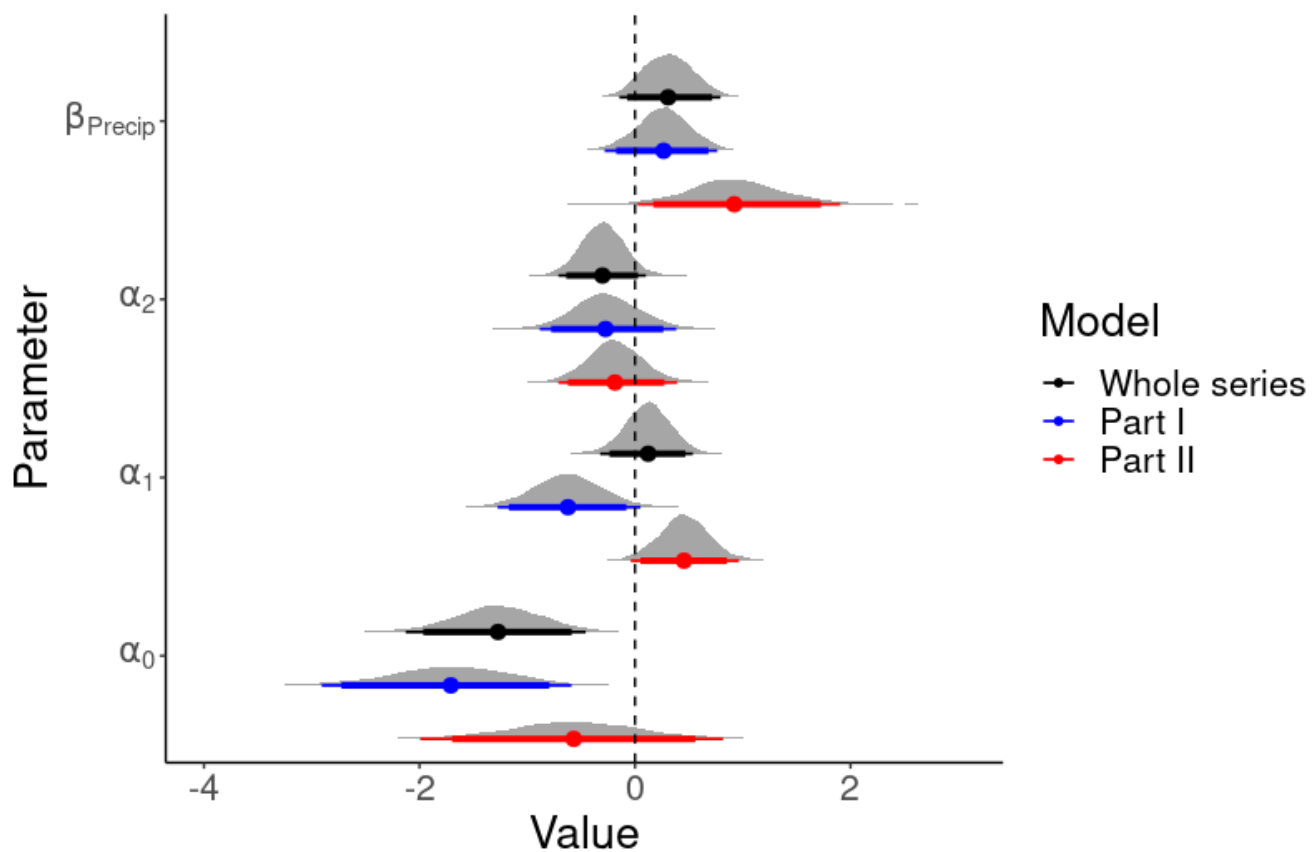
418

**Tables and Figures**



419

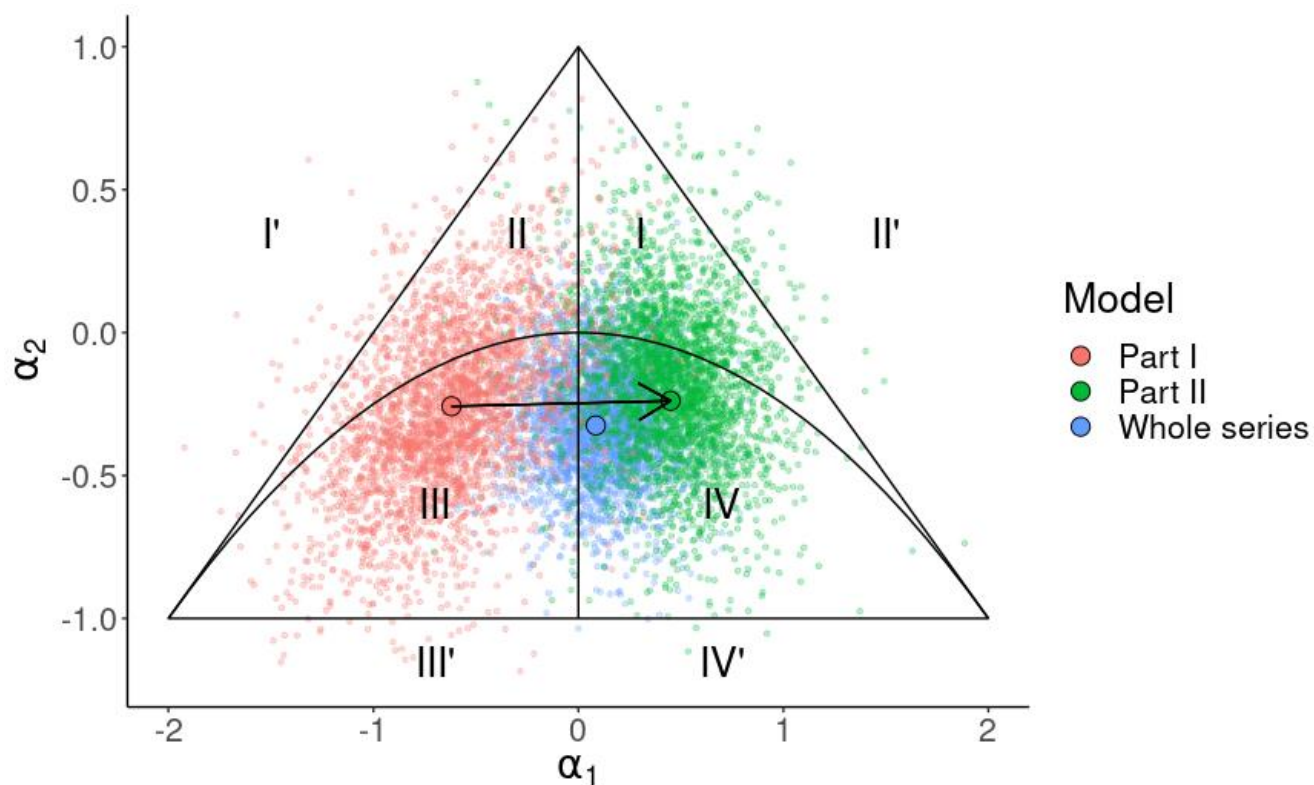
420 **Figure 1:** (a) Time series of caterpillar population counts per 100 m<sup>2</sup> (solid black line) and total annual  
 421 precipitation (dashed blue line); (c) local wavelet transform of caterpillar population counts; (e) local  
 422 wavelet transform of total annual precipitation; (g) wavelet coherence between total annual  
 423 precipitation and caterpillar population counts; (b,d,f) global spectra of (a,c,e); (h) global coherence of  
 424 (g), total annual precipitation and caterpillar population counts. The dashed red line through (a,c,e,g)  
 425 represents the time threshold found in the change-point analysis including precipitation. Caterpillar  
 426 density and spectral period are shown on a log scale; total annual precipitation is shown on the right  
 427 axis. Solid black lines in (c,e,g) delimit regions of significant periodicity or coherence at a 95%  
 428 confidence level from a bootstrap test. Color bars in (c,e,g) show the scale power from low (blue) to  
 429 high (red). Caterpillar wavelet spectrum is shown in (a) with a solid line, and precipitation wavelet  
 430 spectrum is shown with a dashed line. Dashed red lines in (d,f) show the 95% confidence threshold  
 431 from a bootstrap test; peaks to the right of the line represent significant periodicity.



432

433 **Figure 2:** Bayesian posterior 90% and 95% (broad to narrow lines) posterior intervals, and point  
 434 estimates of parameters from Poisson state-space models. Estimates from the whole series are shown  
 435 in black (1986-2019), before the threshold in blue (1986-2004), and after the threshold in red (2004-  
 436 2019).

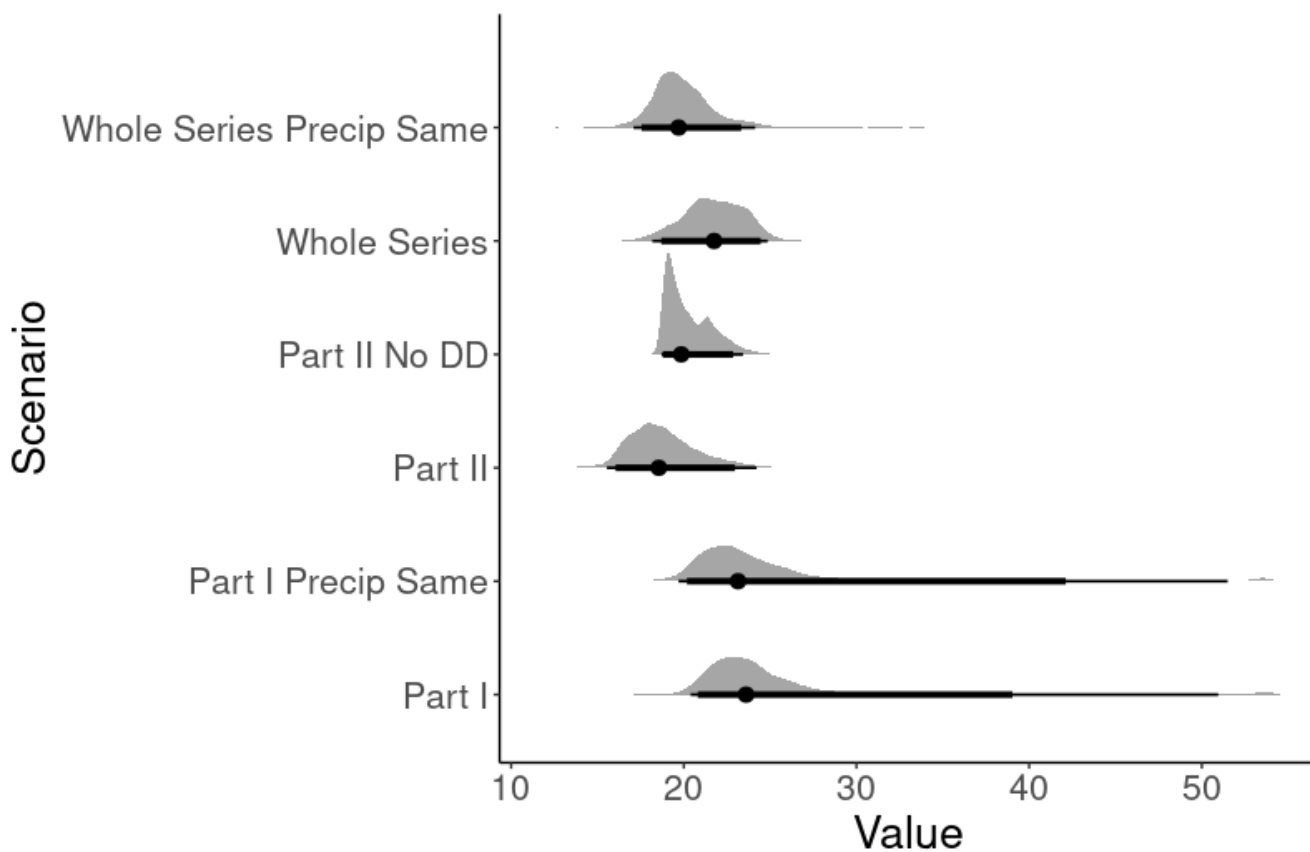
437



438

439 **Figure 3.** Posterior distributions of Bayesian state space models fit to the first part of the series (1986-  
 440 2004), the second part of the series (2004-2019) and the whole series (1986-2019), plotted in the  
 441 Royama parameter plane, showing a shift from type III to type IV dynamics. Open circles represent the  
 442 median of the posterior distributions.

443



444

445 **Figure 4.** Simulation 90% and 95% (broad to narrow lines) posterior intervals and point estimates of  
446 dissimilarity values relative to the true population trajectories, from multiple mechanistic scenarios.

447

448

449

450

451

452

453

454

455

456

457

Model	Period	$\alpha_1$ [P]	$\alpha_2$ [P]	$\beta_1$ [P]
<b>Without Precipitation</b>	<b>1986-2002</b>	-0.56 [P=0.098]	-0.44 [P=0.211]	
	<b>2002-2019</b>	0.35 [P=0.109]	-0.52 [P=0.019]	
<b>With Precipitation</b>	<b>1986-2004</b>	-0.57 [P=0.091]	-0.41 [P=0.209]	-0.01 [P=0.098]
	<b>2004-2019</b>	0.510 [P=0.027]	-0.286 [P=0.218]	1.27 [P=0.026]

458 **Table 1.** Results of change-point analysis, including parameter estimates and P values, from models  
 459 with and without precipitation, before and after change-point thresholds.

460

461

462

463

464

465

466

467

468

469

470

471

472

473

474

475

476

## Supplementary Materials:

Process model	$\Delta$ WAIC	WAIC	$a_0$ [2.5%,97.5%]	$a_1$ [2.5%,97.5%]	$a_2$ [2.5%,97.5%]	$\beta_1$ [2.5%,97.5%]
1986-2019:						
$X_t \sim a_0 + a_1 X_{t-1} + a_2 X_{t-2} + \beta_1 \text{Precip}_{t-1}$	1.38	201.78	-1.28 [-2.13, -0.46]	0.11 [-0.31, 0.53]	-0.30 [-0.70, 0.01]	0.30 [-0.18, 0.78]
$X_t \sim a_0 + a_1 X_{t-1} + \beta_1 \text{Precip}_{t-1}$	1.00	201.40	-1.01 [-1.50, -0.52]	0.03 [-0.29, 0.35]	-	0.36 [0.01, 0.70]
$X_t \sim a_0 + a_1 X_{t-1} + a_2 X_{t-2}$	1.36	201.76	-1.34 [-1.62, -0.52]	0.11 [-0.32, 0.53]	-0.35 [-0.750, 0.05]	-
$X_t \sim a_0 + a_1 X_{t-1}$	0.96	200.76	-1.05 [-1.75, -0.36]	0.05 [-0.37, 0.46]	-	-
$X_t \sim a_0 + \beta_1 \text{Precip}_{t-1}$	<b>0.0*</b>	<b>200.4*</b>	0.05 [-0.28, 0.38]	-	-	0.40 [0.018, 1.04]
$X_t \sim a_0$	0.13	200.53	0.04 [-0.59, 0.68]	-	-	-
1986-2004:						
$X_t \sim a_0 + a_1 X_{t-1} + a_2 X_{t-2} + \beta_1 \text{Precip}_{t-1}$	0.52	117.37	-1.74 [-2.92, -0.61]	-0.63 [-1.29, 0.03]	-0.27 [-0.89, 0.34]	0.26 [-0.27, 0.78]
$X_t \sim a_0 + a_1 X_{t-1} + \beta_1 \text{Precip}_{t-1}$	0.57	117.41	-1.48 [-2.17, -0.81]	-0.57 [-1.10, -0.03]	-	0.27 [-0.24, 0.77]
$X_t \sim a_0 + a_1 X_{t-1} + a_2 X_{t-2}$	<b>0.00*</b>	<b>116.85*</b>	-1.68 [-2.87, -0.55]	-0.57 [-1.23, 0.08]	-0.26 [-0.90, 0.38]	-
$X_t \sim a_0 + a_1 X_{t-1}$	0.11	116.96	-1.46 [-2.17, -0.78]	-0.53 [-1.06, -0.03]	-	-
$X_t \sim a_0 + \beta_1 \text{Precip}_{t-1}$	0.72	117.57	-0.17 [-1.06, 0.72]	-	-	0.09 [-0.81, 0.98]
$X_t \sim a_0$	0.81	117.66	-0.18 [-1.03, 0.68]	-	-	-
2004-2019:						
$X_t \sim a_0 + a_1 X_{t-1} + a_2 X_{t-2} + \beta_1 \text{Precip}_{t-1}$	<b>0.00*</b>	<b>86.32*</b>	-0.58 [-1.98, 0.79]	0.46 [-0.05, 0.96]	-0.19 [-0.74, 0.36]	0.92 [0.01, 1.87]
$X_t \sim a_0 + a_1 X_{t-1} + \beta_1 \text{Precip}_{t-1}$	1.65	87.97	-0.22 [-1.50, 0.96]	0.60 [0.01, 1.15]	-	1.14 [0.24, 1.15]
$X_t \sim a_0 + a_1 X_{t-1} + a_2 X_{t-2}$	0.53	86.85	-1.30 [-2.78, 0.06]	0.340 [-0.23, 0.89]	-0.46 [-1.02, 0.01]	-
$X_t \sim a_0 + a_1 X_{t-1}$	1.13	87.45	-0.72 [-2.27, 0.71]	0.35 [-0.32, 1.00]	-	-
$X_t \sim a_0 + \beta_1 \text{Precip}_{t-1}$	1.76	88.08	0.45 [-0.35, 1.23]	-	-	1.37 [0.45, 2.28]
$X_t \sim a_0$	1.02	87.34	0.33 [-0.79, 1.46]	-	-	-

478 **Table S1.** Results of Bayesian state space time series model selection, by period of series analyzed and  
479 the process model structure. Change in WAIC relative to the best model ( $\Delta$ WAIC), WAIC, and parameter  
480 estimates with 95% posterior intervals are shown. WAIC values within  $\Delta$ WAIC <2 of the best model are  
481 bolded, and the best model is bolded. Parameter estimates the 95% posterior intervals which do not  
482 overlap zero are bolded as well.

483

484

485

486

487

488

489

490

491

492

493

494

495

Series	Original	Whole series, fitted $\beta_1$	Part I, fitted $\beta_1$	Part II, fitted $\beta_1$	Whole series, equal $\beta_1$	Part I, equal $\beta_1$	Part II, equal $\beta_1$	Part II, no $a_1, a_2$
Original	0							
Whole series, fitted $\beta_1$	22.14	0						
Part I, fitted $\beta_1$	22.80	10.38	0					
Part II, fitted $\beta_1$	<b>17.55</b>	10.17	16.51	0				
Whole series, equal $\beta_1$	18.54	11.87	15.89	11.30	0			
Part I, equal $\beta_1$	21.42	19.56	18.19	16.44	17.58	0		
Part II, equal $\beta_1$	<b>17.55</b>	10.17	16.51	0	6.22	16.5	0	
Part II, no $a_1, a_2$	19.72	12.73	15.84	6.52	6.37	13.75	6.67	0

496 **Table S2.** Dissimilarity matrix of simulated time series from fitted state-space process models, with  
 497 original fitted parameters and with effect of precipitation set to equal at the highest fitted value  
 498 ( $\beta_1=0.922$ ), all compared with the original observed time series. Lowest dissimilarities relative to the  
 499 original series for simulations with fitted or equal  $\beta_1$  values are bolded.

500

501

502

503

504

505

506

507

508

509

510



511

<i>Series</i>	<i>Parameter</i>	$a_0$	$a_1$	$a_2$	$\beta_1$
<i>Whole Series</i>	$a_0$	1			
	$a_1$	0.584	1		
	$a_2$	0.551	-0.014	1	
	$\beta_1$	0.109	0.001	0.157	1
<i>Part I</i>	$a_0$	1			
	$a_1$	0.744	1		
	$a_2$	0.739	0.465	1	
	$\beta_1$	-0.116	-0.214	-0.049	1
<i>Part II</i>	$a_0$	1			
	$a_1$	0.491	1		
	$a_2$	0.628	-0.116	1	
	$\beta_1$	0.501	0.207	0.520	1

512 **Table S3.** Cross-correlation matrices between parameter estimates from MCMC chains the full state

513 space model from the whole series, part I and part II.

514

515

516

517

518

519

520

521

522

523

524

525

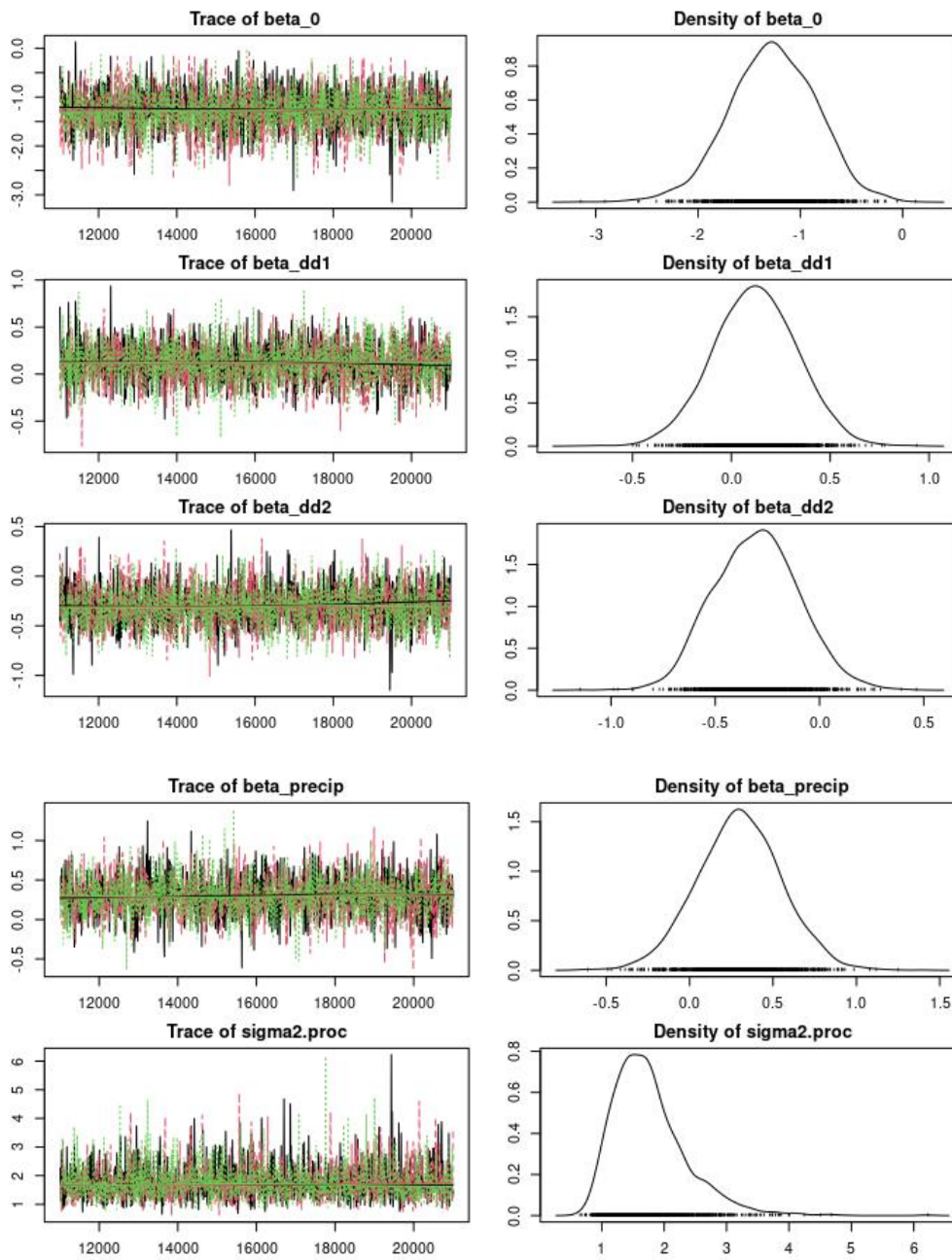
526

Year	Caterpillar count	Lupine area (m <sup>2</sup> )
1986	490	400
1987	8	101.35
1988	12	53.75
1989	25	51
1990	48	61.24
1991	7	41.5
1992	71	63.75
1993	5	29.25
1994	16	59.75
1995	15	20.5
1996	40	18.05
1997	2	20
1998	8	20
1999	12	18
2000	8	38
2001	40	31.5
2002	18	63
2003	52	45.5
2004	3	64.5
2005	0	57.5
2006	2	34
2007	51	35
2008	40	32.5
2009	9	62
2010	2	51.5
2011	23	27.75
2012	27	15.5
2013	2	40
2014	8	33
2015	3	29.5
2016	13	38
2017	7	30
2018	22	21
2019	131	26

527 **Table S4.** Year, caterpillar counts, and area of lupine surveyed.

528

529

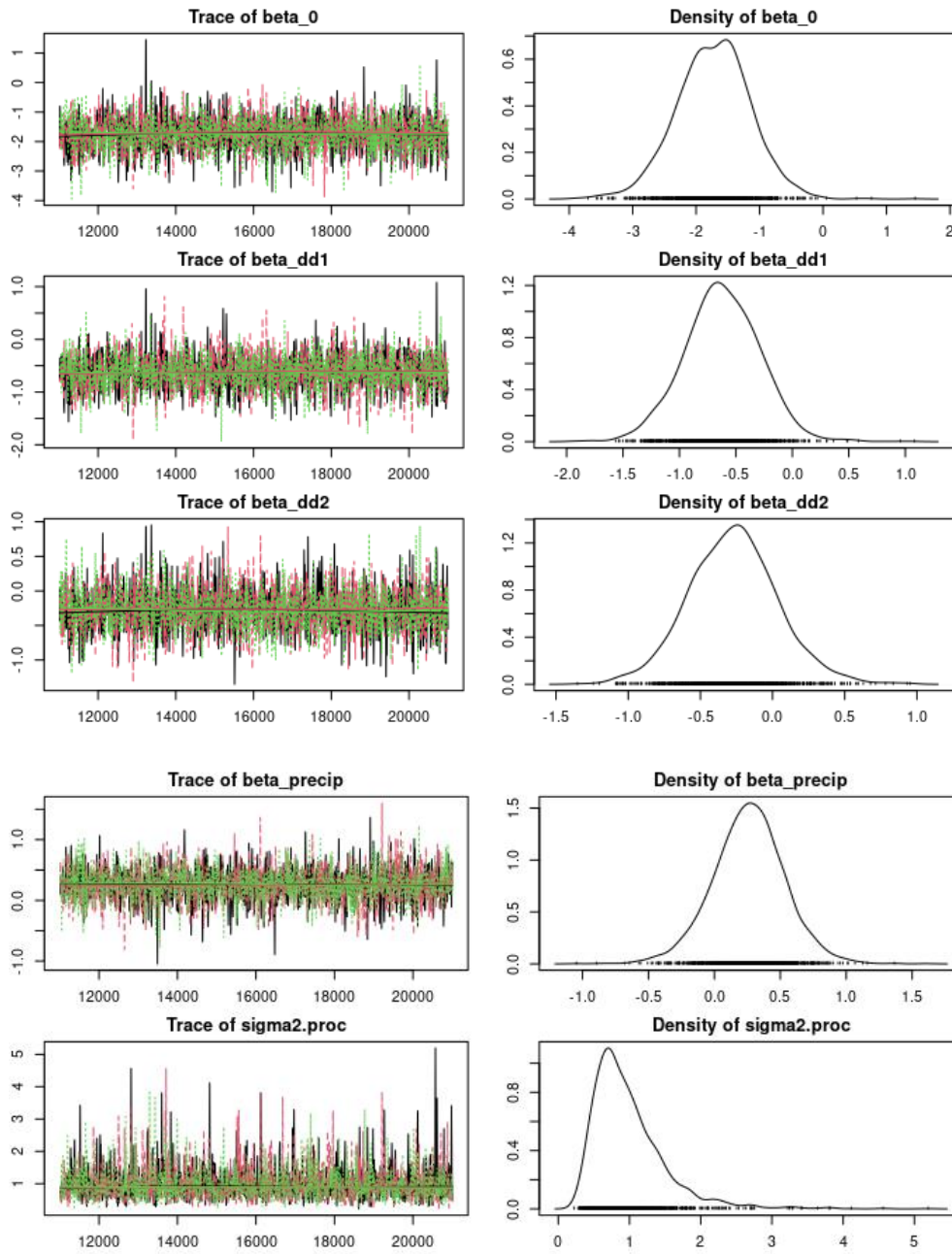


530

531

532 **Figure S1.** MCMC trace and density plot of parameters from full state space model of the whole  
533 series.

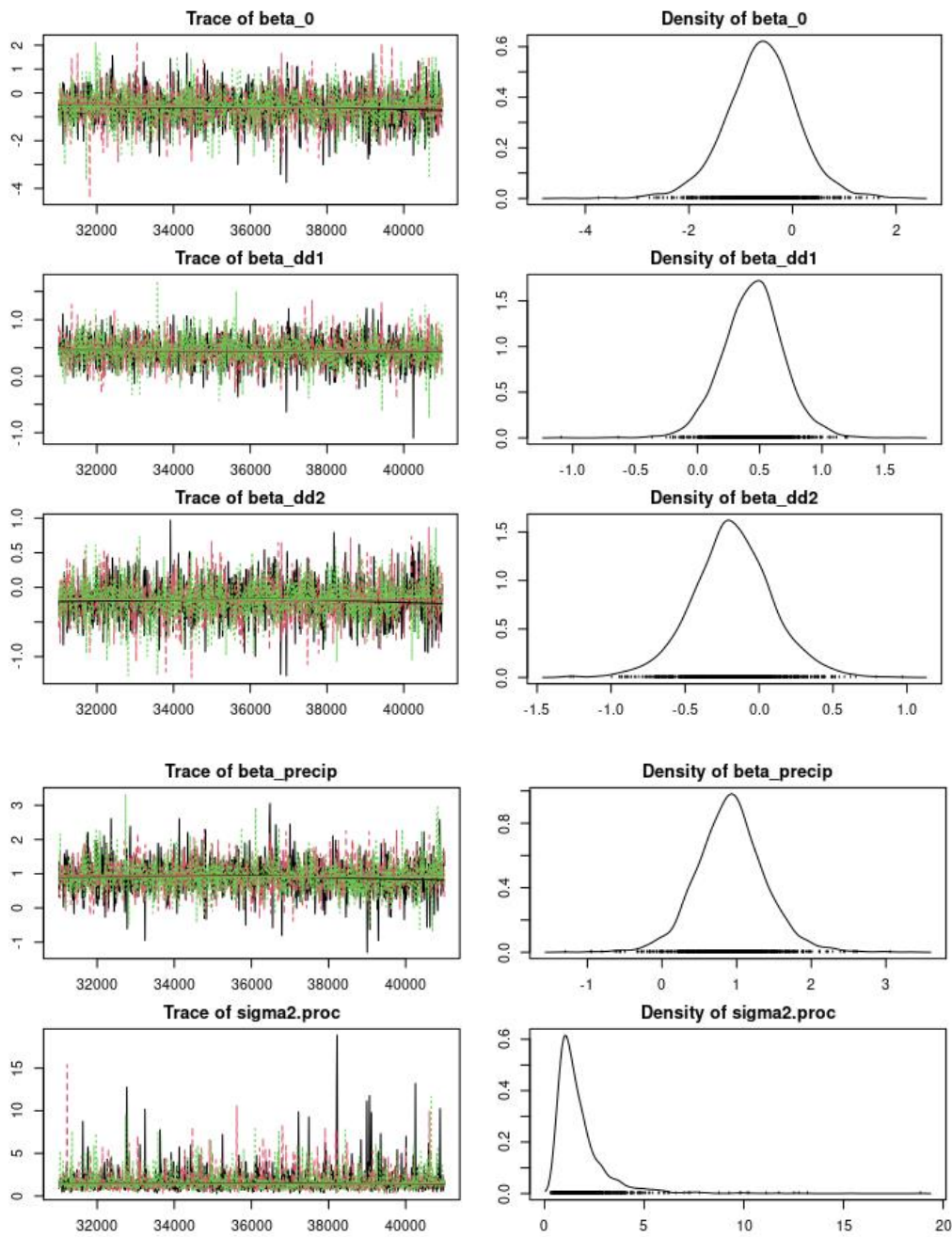
534



535

536

537 **Figure S2.** MCMC trace and density plots of parameters from full state space model of the first part of  
538 the series.

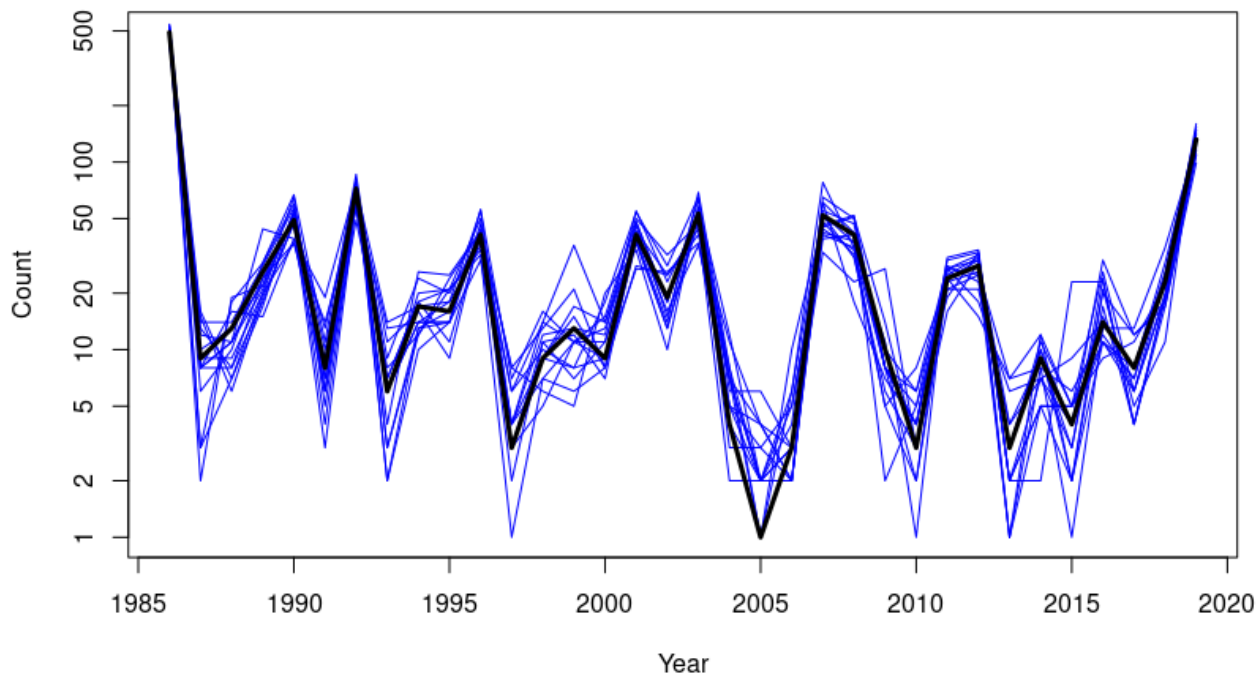


539

540

541 **Figure S3.** MCMC trace and density plots of parameters from full state space model of the second part  
542 of the series.

543

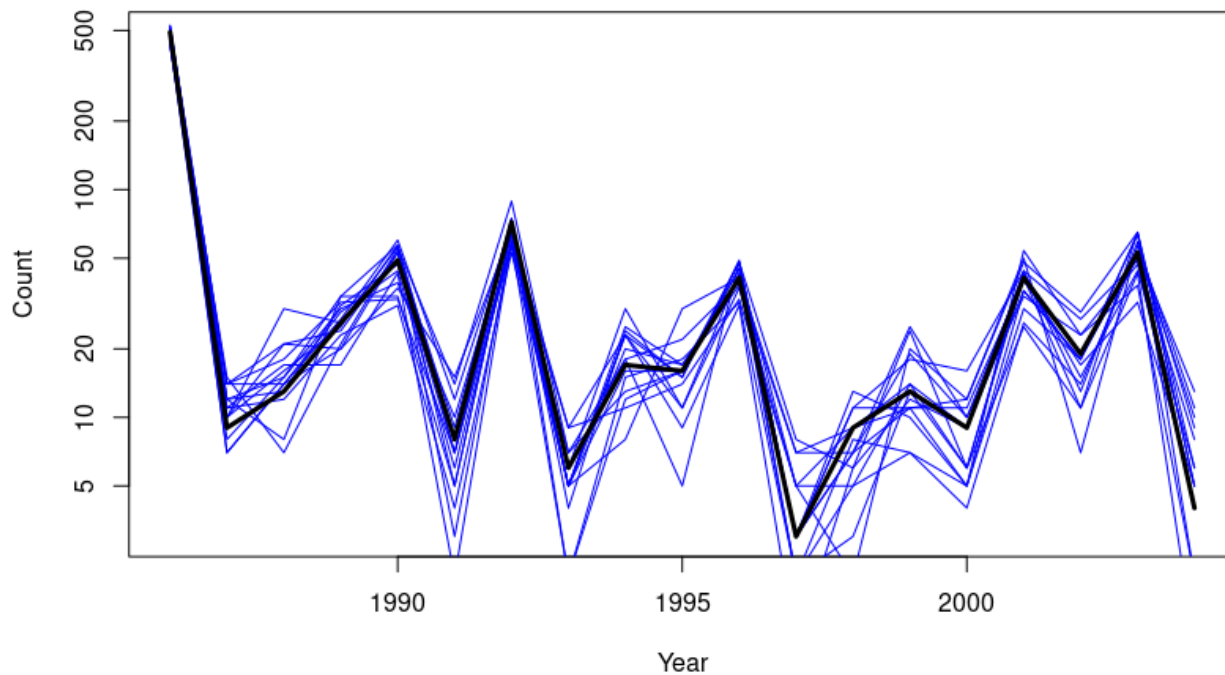


544 **Figure S4.** Posterior predictive simulations from full state space model of the whole series, with the  
545 observed trajectory in black, and one-step ahead simulations from 15 draws from the posterior in blue.

546

547

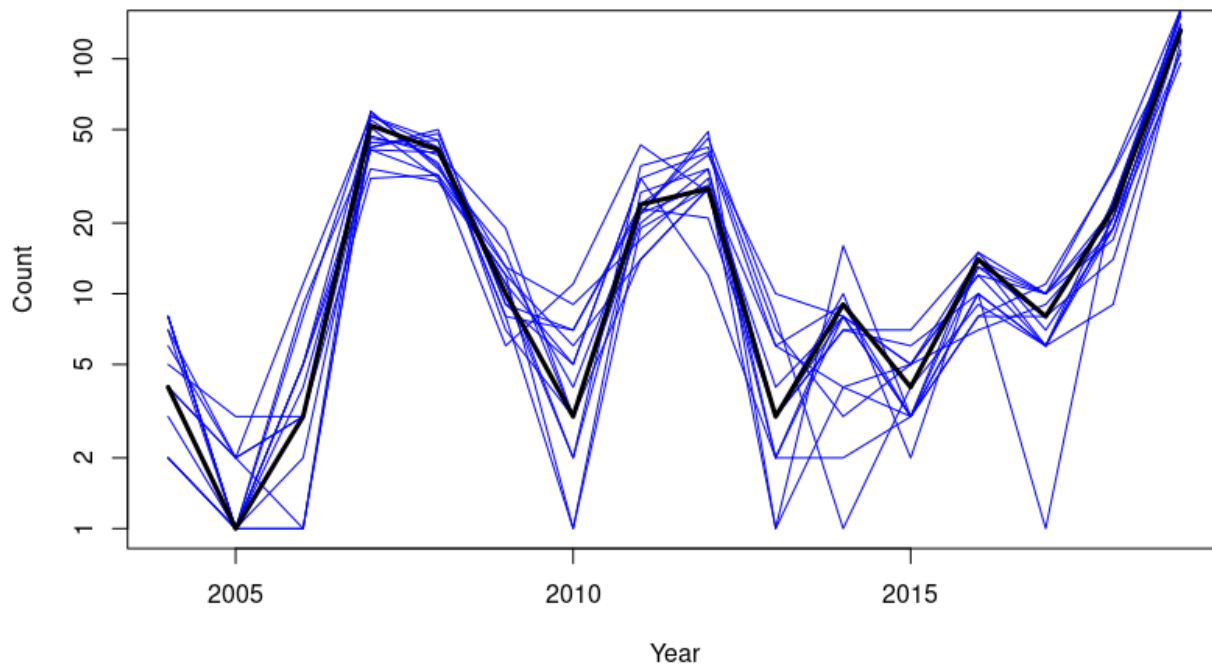
548



549

550 **Figure S5.** Posterior predictive simulations from full state space model of the first part of the series,  
551 with the observed trajectory in black, and one-step ahead simulations from 15 draws from the posterior  
552 in blue.

553



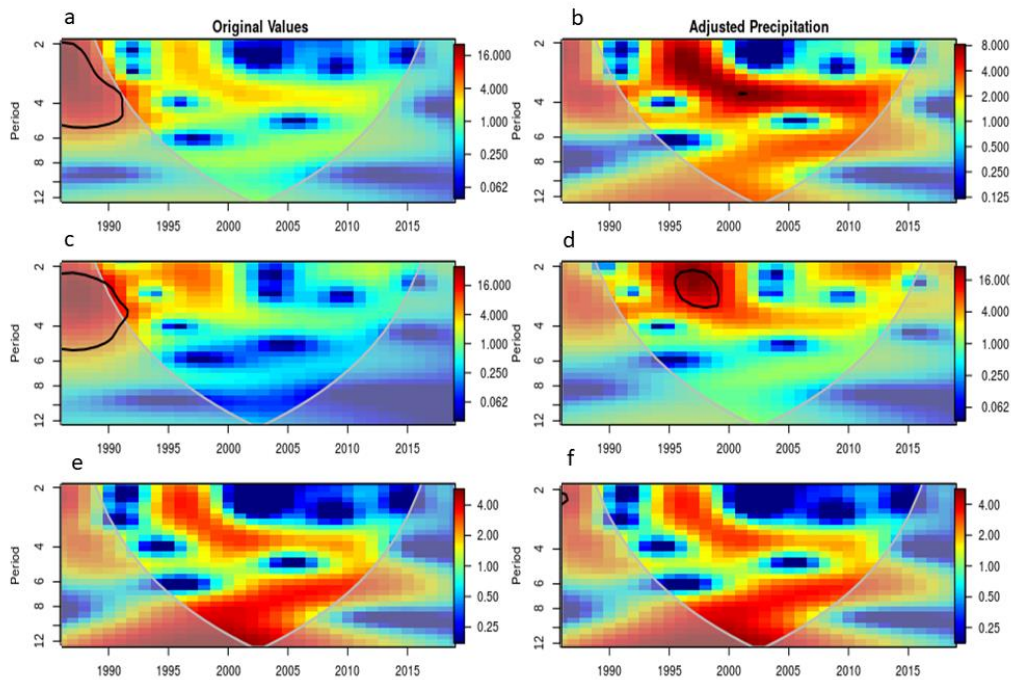
554

555 **Figure S6.** Posterior predictive simulations from full state space model of the second part of the series,  
556 with the observed trajectory in black, and one-step ahead simulations from 15 draws from the posterior  
557 in blue.

558

559





560

561 **Figure S7.** Wavelet transformations of simulated trajectories from fitted state-space process models.

562 (a,c,e) Simulations using original fitted values, and (b,d,f) using fitted values for density-dependence

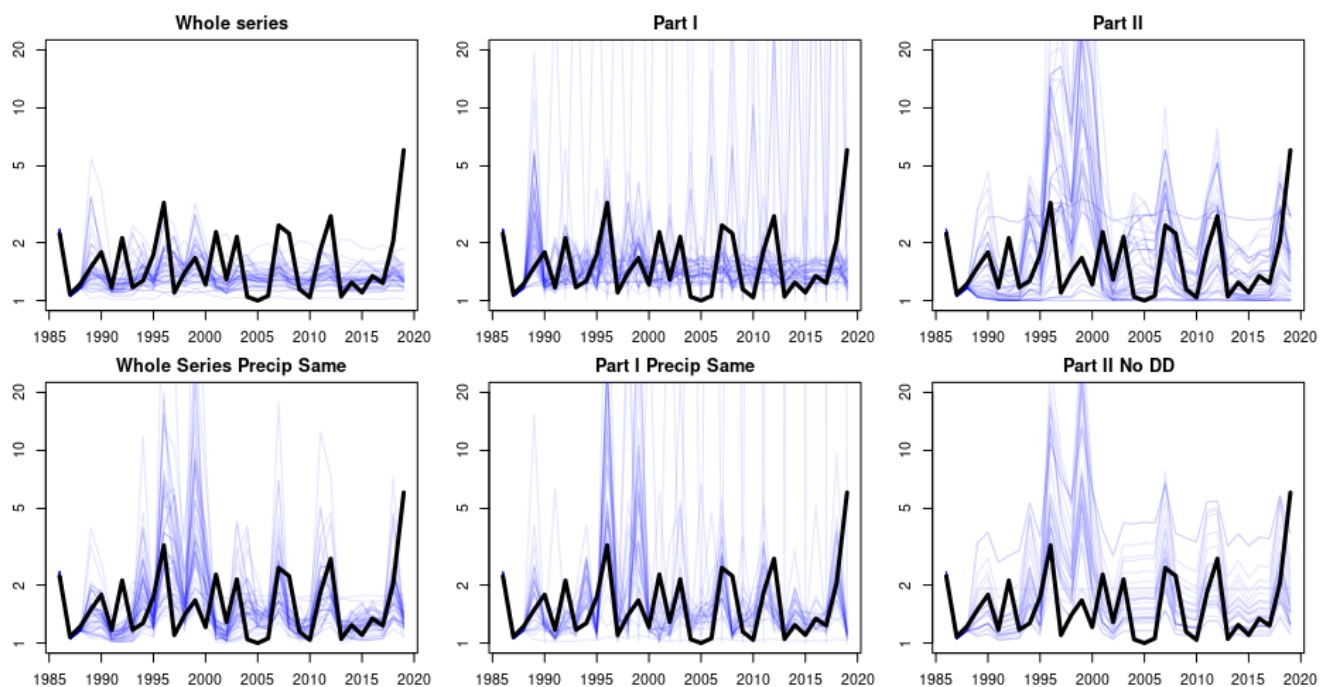
563 but the highest values for precipitation. (a,b) Models of the whole series, (c,d) before the threshold, and

564 (e,f) after the threshold. Black lines delimit regions of significant periodicity or coherence at a 95%

565 confidence level from a bootstrap test. Color bars show the scale for power from low (blue) to high

566 (red).

567



568

569

570

571

572

573

**Figure S8.** Fifty simulation trajectory results (blue) vs. true population trajectory (black) from simulation based on the models (left to right, top to bottom) from the whole series, part I, part II, the whole series with precipitation values from part II, part I with precipitation values from part II, and part II without density dependence.